

University of Windsor

Scholarship at UWindsor

Great Lakes Institute for Environmental
Research Publications

Great Lakes Institute for Environmental
Research

2016

Is it appropriate to composite fish samples for mercury trend monitoring and consumption advisories?

Ken G. Drouillard

Nilima Gandhi
University of Windsor


Satyendra P. Bhavsar
Ontario Ministry of Environment

Sarah B. Gewurtz
University of Windsor

George B. Arhonditsis
University of Toronto - Erindale College

See next page for additional authors

Follow this and additional works at: <https://scholar.uwindsor.ca/glierpub>

 Part of the [Biochemistry, Biophysics, and Structural Biology Commons](#), and the [Physical Sciences and Mathematics Commons](#)

Recommended Citation

Drouillard, Ken G.; Gandhi, Nilima; Bhavsar, Satyendra P.; Gewurtz, Sarah B.; Arhonditsis, George B.; and Petro, Steve. (2016). Is it appropriate to composite fish samples for mercury trend monitoring and consumption advisories?. *Environmental Science and Technology*.
<https://scholar.uwindsor.ca/glierpub/17>

This Article is brought to you for free and open access by the Great Lakes Institute for Environmental Research at Scholarship at UWindsor. It has been accepted for inclusion in Great Lakes Institute for Environmental Research Publications by an authorized administrator of Scholarship at UWindsor. For more information, please contact scholarship@uwindsor.ca.

Authors

Ken G. Drouillard, Nilima Gandhi, Satyendra P. Bhavsar, Sarah B. Gewurtz, George B. Arhonditsis, and Steve Petro

Is it appropriate to composite fish samples for mercury trend monitoring and consumption advisories?

Nilima Gandhi^a, *Satyendra P. Bhavsar*^{a,b,c,*}, *Sarah B. Gewurtz*^c, *Ken G. Drouillard*^c, *George
B. Arhonditsis*^a, *Steve Petro*^b

^a University of Toronto, Toronto, ON, Canada M1C 1A4

^b Ontario Ministry of the Environment and Climate Change, Toronto, ON, Canada, M9P 3V6

^c University of Windsor, 401 Sunset Avenue, Windsor, ON, Canada, N9B 3P4

* Corresponding author Tel: 1-416-327-5863; fax: 1-416-327-6519.

E-mail addresses: s.bhavsar@utoronto.ca or satyendra.bhavsar@ontario.ca (S.P. Bhavsar).

Abstract

1 Monitoring mercury levels in fish can be costly because variation by space, time, and fish
2 type/size needs to be captured. Here, we explored if compositing fish samples to decrease
3 analytical costs would reduce the effectiveness of the monitoring objectives. Six compositing
4 methods were evaluated by applying them to an existing extensive dataset and examining their
5 performance in reproducing the fish consumption advisories and temporal trends. The methods
6 resulted in varying amount (average 34-72%) of reductions in samples, but all (except one)
7 reproduced advisories very well (96-97% of the advisories did not change or were one category
8 more restrictive compared to analysis of individual samples). Similarly, the methods performed
9 reasonably well in recreating temporal trends, especially when longer-term and frequent
10 measurements were considered. The results indicate that compositing samples within 5 cm fish
11 size bins or retaining the largest/smallest individuals and compositing in-between samples in
12 batches of 5 with decreasing fish size would be the best approaches. Based on the literature, the
13 findings from this study are applicable to fillet, muscle plug and whole fish mercury monitoring
14 studies. The compositing methods may also be suitable for monitoring Persistent Organic
15 Pollutants (POPs) in fish. Overall, compositing fish samples for mercury monitoring could result
16 in a substantial savings (approximately 60% of the analytical cost) and should be considered in
17 fish mercury monitoring, especially in long-term programs or when study cost is a concern.

18

19 *Keywords:* Mercury Hg; compositing/pooling; fish; monitoring; advisories; sensitive population

Graphical Abstract



Highlights

- We test if compositing fish to decrease costs biases Hg monitoring
- Six compositing methods were assessed using an extensive dataset from Ontario
- Five methods reproduced advisories very well (96-97% same or more stringent)
- Methods performed well in recreating temporal trends, especially for longer-terms
- Methods resulted in average 34-72% reductions in samples and should be considered

21 **1. Introduction**

22 Mercury is a contaminant of global concern (UNEP, 2013a). Virtually every fish in North
23 America, and possibly worldwide, contains mercury (Stahl et al., 2009; Depew et al., 2013;
24 Evers et al., 2013). Consumption of fish is generally a dominant route of human exposure to
25 mercury (UNEP/WHO, 2008). Mercury is responsible for the most number of restrictive fish
26 consumption advisories, at least in North America (e.g., USEPA, 2013a, b; OMOECC, 2015).
27 Due to spatial variation in fish mercury levels, location-specific advisories are typically provided
28 (e.g., USEPA, 2013a; OMOECC, 2015). Since mercury levels vary by fish species and size
29 (Gewurtz et al., 2011b), monitoring efforts to issue fish consumption advisories and track long-
30 term changes require collection and analysis of a variety of fish spanning their natural size range
31 (USEPA, 2013b). As a result, the total number of annual samples required to adequately monitor
32 fish mercury levels for numerous locations can range from hundreds to tens of thousands.

33 Due to analytical costs, most contaminant studies limit sample size by reducing the fish
34 species monitored, replication of samples, sampling frequency and/or study period; however,
35 these options are generally not suitable for agencies that rely on the data for long-term trend
36 monitoring and issuing of fish consumption advisories aimed at protecting human health
37 (Gewurtz et al., 2011a). Further, Article 19 of the recently formulated Minamata Convention on
38 Mercury requires parties to develop and improve geographically representative mercury
39 monitoring in environmental media, including fish (UNEP, 2013b). In less than a decade,
40 monitoring data will be called upon to assist in the implementation and evaluation of the
41 convention, which emphasizes the importance of improving monitoring efforts to optimize both
42 the quality of the programs as well as costs.

43 To decrease program costs, combining multiple temporally or spatially discrete samples,
44 widely known as composites, has been suggested as an effective alternative to chemical analysis
45 on individual samples (USEPA, 2002; Gewurtz et al., 2011a). In addition to substantially
46 reducing analytical cost, the data collected through compositing samples can provide wider
47 temporal and spatial coverage without increasing the sample count. The analysis of data may
48 give more representative estimates of mean concentrations than can the same number of discrete
49 sampling, albeit at the cost of variability in the observations (USEPA, 2002).

50 There are several potential approaches to compositing fish contaminant monitoring
51 samples that incorporate different dimensions of the study, such as time (within/across years),
52 location, fish species, and fish size. The optimal compositing approach would be one that
53 reduces the total number of samples for analysis without compromising the objectives of the
54 monitoring program. In addition, the composite method chosen should follow assumptions that
55 correspond to the statistical analysis that is ultimately applied to the data. Several studies have
56 used compositing as a part of their designs for both organic and inorganic contaminants in all
57 media including biota (Rajagopal and Williams, 1989; Turle and Collins, 1992; Blomqvist, 2001;
58 Braune and Noble, 2009; Gewurtz et al., 2011a). However, to our knowledge, a comprehensive
59 study investigating the effectiveness of various compositing approaches for monitoring mercury
60 in fish is lacking in the literature, especially for programs designed to generate fish consumption
61 advice, where variability and the presence of outliers can affect overall risk (Gewurtz et al.,
62 2011a).

63 In this study, we evaluate six methods of compositing fish samples by examining their
64 performance if they would have been utilized instead of collecting >220,000 individual mercury
65 measurements for >3000 locations by the Province of Ontario, Canada over nearly 50 years. The

66 effectiveness of the composite methods was evaluated by comparing the fish consumption
67 advisories and temporal trends from individual measurements (current sampling design) with
68 estimated composite values, calculated by averaging the individual measurements included in
69 each composite. The findings of the study determines whether a compositing method can
70 effectively minimize costs for regular, long term, large scale monitoring programs and set
71 advisories for fish consumption.

72

73 **2. Methods**

74 ***2.1 Compositing Methods***

75 Fish mercury levels vary by species and size, and can change seasonally as well as over
76 time under the influence of a variety of internal and external factors, such as bioenergetics and
77 ambient water chemistry (Bhavsar et al., 2010; Azim et al., 2011; Gewurtz et al., 2011a; Stern et
78 al., 2012; Greenfield et al., 2013). As such, we opted to group species-specific samples collected
79 during the same sampling event within the composites.

80 There is a well-known relationship between mercury concentrations and fish size that is
81 typically described by the power-series regression (Gewurtz et al., 2011b). As such, similar
82 sized samples could be considered for creating a composite sample. However, the resultant fish
83 size range (i.e., maximum-minimum fish lengths) would likely be less than the regular,
84 individual measurements. This could result in trimming of a regression at the extreme ends, and
85 thereby loss of advisories for certain fish sizes. Alternatively, if one or two of the largest and/or
86 smallest individuals are retained with all other samples being composited, then the fish size

87 range could be captured, and a power series regression between fish length and composited
88 mercury concentrations might be improved.

89 Compositing of 3, 5, 7, 10 or more samples have been used in many studies (Hites et al.,
90 2004; Carlson and Swackhamer, 2006; French et al., 2011; Pantazopoulos et al., 2013). Since a
91 collection of about 20 fish samples per species and sampling event over a possible maximum
92 size range is generally considered a preferred method for mercury monitoring (e.g., Gewurtz et
93 al., 2011a), compositing more than 5 samples (i.e., having less than four composites), may not be
94 sufficient for characterizing the fish size/mercury relationships. Alternatively, compositing
95 samples within a narrow size range (e.g., 35-40 cm, 40-45 cm and so on) regardless of the
96 number of samples within that size range may be appropriate as the impact on the fish
97 size/mercury relationship would likely be minimal.

98 Based on the above notes, we considered six compositing methods: (1) composite
99 samples in batches of five in the order of decreasing fish size (Figure 1a,b), (2) retain individual
100 samples for the largest and smallest fish and composite samples in between in batches of five in
101 order of decreasing fish size (Figure 1a,c), (3) retain the two largest and smallest individual
102 samples and composite the samples in between in batches of five in order of decreasing fish size
103 (Figure 1a,d), (4) retain the largest and smallest individual samples and composite the samples in
104 between in batches of three in order of decreasing fish size (Figure 1a,e), (5) retain the two
105 largest and smallest individual samples and composite the samples in between in batches of three
106 in the order of decreasing fish size (Figure 1a,f), and (6) composite samples within a 5 cm size
107 range (Figure 1a,g).

108 **2.2 Data Source**

109 The above described compositing methods were evaluated by simulating composite data
110 from the individual fish measurements, assuming that the same mass of each fish is added to the
111 composite. For this purpose, we used an extensive and consistent fish mercury dataset
112 comprising 223,318 individual, widely varying measurements for skinless, boneless dorsal fillets
113 of >10 cm fish of 66 fish species (Table S1) collected by the Ontario Ministry of the
114 Environment and Climate Change (OMOECC), Canada in partnership with the Ontario Ministry
115 of Natural Resources and Forestry and other agencies over nearly 50 years (1967-2014) from
116 >3000 locations in the Province of Ontario, Canada, that spans 41° to 56° N and 74° to 95° W
117 (Figure S1). The samples were analysed for total mercury using acid digestion and cold vapor
118 flameless atomic absorption spectroscopy as described in detail by Bhavsar et al. (2010). The
119 dataset contained 16,900 species/location/year combinations for 6,440 sampling events
120 (location/year) and varied widely (1 to 274) in the number of individual samples for a species in
121 a sampling event (species/location/year) (Figure S2).

122 ***2.3 Statistical analysis***

123 The performance of each composite method in comparison to the regular, individual
124 measurements was evaluated based on its accuracy in reproducing the fish consumption
125 advisories as well as the direction and magnitude of the long-term temporal trends. As illustrated
126 in Figure S3, a power series regression was conducted for each of 16,900 species/location/year-
127 specific sampling events using the regular, individual measurements as well as the composite
128 values calculated using the six methods considered in this study. Using these total 118,300
129 power series regressions (i.e., 16,900 x 7), fish mercury levels were calculated at 5 cm intervals
130 for the available size range in each species-specific sampling event (Figure S3). These mercury
131 concentrations were used in calculating fish consumption advisories using the benchmarks for

132 the general population and sensitive population (children and women of child-bearing age),
133 which is the standard method used by the Province of Ontario, Canada (Table S2, Figure S3).
134 Advisories for each 5 cm interval calculated using the six composite methods were compared
135 with those from the regular, individual measurements (Table S4), and classified into three
136 categories: 1) same, 2) more restrictive, and 3) less restrictive.

137 For a comparison of temporal trend analyses from the regular and composite methods,
138 rates of changes in fish mercury levels ($\mu\text{g/g}$ decade) were calculated using the slope of the
139 linear relationship between year and mercury concentration standardised to a fish length. Since
140 the purpose is to compare rates from the regular and composite methods, appropriateness of a
141 linear regression is essentially a moot point (Azim et al., 2011). Since a temporal trend analysis
142 is typically conducted on a suitable indicator species with good monitoring data, four species,
143 namely Lake Trout (*Salvelinus namaycush*), Walleye (*Sander vitreus*), Northern Pike (*Esox*
144 *Lucius*) and Smallmouth Bass (*Micropterus dolomieu*), were considered. Mercury
145 concentrations standardized to 50 cm fish size were used. The standardization was conducted
146 using a power series regression $y = a x^b$, where y is concentration in $\mu\text{g/g}$, x is fish length in cm,
147 and a and b are regression coefficients. The number of temporal trend rate estimates was
148 maximized by considering every combination of the start and end years as illustrated in Figure
149 S4. In total, 83,664 rates of fish mercury changes were calculated. All statistical analyses were
150 conducted in either Excel 2010 or R-3.2.0 for Windows™ (R Core Development Team, 2015).

151

152 **3. Results**

153 **3.1 Reductions in samples**

154 The composite method 1 resulted in the highest (average/median 72/78%) reduction in
155 number of samples to be analysed for mercury (Figure 2). The composite methods 2 and 3
156 required retention of one and two extreme sized individual samples, respectively. As such, the
157 reductions in number of samples were less (method 2: 54/64%; method 3: 40/50%; Figure 2).
158 The methods 4 and 5 required compositing samples in the batches of 3, compared to 5 for the
159 methods 2 and 3. As a result, reductions in the number of samples by implementing the methods
160 4 and 5 were less (method 4: 45/53%; method 5: 34/42%; Figure 2). Although the composite
161 method 6 resulted in more variable (0-98%) reductions in the samples because of its dependence
162 on number of samples in 5 cm fish size bins, overall reductions were similar to the method 2
163 (55/60%; Figures 2, S5).

164 ***3.2 Performance in reproducing advisories***

165 Seven sets of fish consumption advisories (regular plus six composite methods) were
166 calculated for each sampling event (species/location/year) as illustrated in Figure S3, and
167 compared as shown in Table S3. The resultant fish size ranges (minimum to maximum length)
168 for the composite method 1 were lower than from the regular, individual measurements for many
169 sampling events. In addition, method 1 produced one composite for each of 3,681 sampling
170 events with ≤ 5 samples (Figure S2), resulting in no power series regression for an advisory
171 calculation. Therefore, about 35% of the advisories from method 1 were missing (Figure 3,
172 Table S3).

173 The advisories were calculated using power series regressions on fish size vs mercury
174 concentrations for each sampling event (location/year/species). The statistical significance of the
175 regressions was evaluated on the basis of their p-values. Since the composites were aimed at

176 reducing the sample size, which is generally positively related to a p-value of a regression, it was
177 not surprising to observe lower statistical significance for regressions from a composite method
178 that produced a greater reduction in sample sizes (Figures 2, S6).

179 Overall, advisories for the general population from the methods 2 to 6 were largely (85-
180 91%) similar to those from the regular, individual measurements (Figure 3, Table S4a). About
181 6-11% of the advisories were more restrictive, mostly by only one advisory category (Figure 3,
182 Table S4a). Only 3-4% of the advisories were less restrictive, again mostly by only one advisory
183 category (Figure 3, Table S4a). The results for the sensitive population advisories were even
184 better (similar: 88-93%; more restrictive 5-9%; less restrictive 2-3%; Figure 3, Table S4b).

185 The increasingly fewer reductions in the number of samples from the composite methods
186 2 to 5 only marginally improved reproduction of the advisories (Figure 3). The performance of
187 the method 6 was similar to the method 4 and overall second best among the methods (Figure 3,
188 Table S4). Based on the reductions in the number of samples and performances in reproducing
189 the advisories, we focus further analysis and the following discussions on results for the general
190 population using the methods 2 and 6.

191 Next we examined if there was a pattern in the underestimation of mercury
192 concentrations and thereby less restrictive advisories from the composite methods that could be
193 linked to sample size, species, fish size class, and/or level of mercury. As shown in Tables S5-
194 S8, individually these four factors had minimal impact on the performance of the composite
195 methods 2 and 6. The only exception was that increasing fish size worsened the performance of
196 method 2, with relatively more cases of less restrictive advisories for large size categories within
197 individual species (Table S9). Nevertheless, there were only 3-4 combinations of species/size

198 for which the total number of advisories were >100 and >10% of the advisories were less
199 restrictive (Table S9). Similarly, there was no fish species-specific mercury concentration that
200 substantially affected the performance of the composite methods 2 and 6 (Table S10).

201 ***3.3 Performance in reproducing temporal trends***

202 In this assessment, we examined if the nature of the mercury versus time slopes from the
203 composite methods corresponded with the regular method. The composite methods resulted in
204 the same temporal trends as observed for the individual samples in most (90-94%) cases (Figure
205 S7). The performances of the composite methods improved from 90-94% to 94-96% when cases
206 with a minimum time span of 15 years and 5 sampling years were considered, and to 95-97%
207 when cases with a minimum time span of 15 years and 10 sampling years were considered
208 (Figure S7).

209 For a majority (72-82%) of the cases, the rates of changes in fish mercury levels from the
210 composite methods were within a factor of two of the corresponding rates from the regular
211 method (Figure S8). Approximately 81-88% of the rates were within a factor of three (Figure
212 S8). When cases with a minimum time span of 15 years and 5 sampling years were considered,
213 the percentages of cases improved to 81-88% for within a factor of two and 88-92% for within a
214 factor of three (Figure S8). The corresponding results for cases with a minimum time span of 15
215 years and 10 sampling years were better at 83-90% and 89-93%, respectively (Figure S8).

216 The performance of the composite methods in reproducing the rates of changes was also
217 evaluated for each of the four selected fish species. All composite methods provided the same
218 temporal trends for a majority (83-95%) of the cases for all species (Figure S9). When cases
219 with a minimum time span of 15 years and 10 sampling years were considered, the percentages

220 of cases improved to 97-100% for Lake Trout, Northern Pike and Walleye, and 86-90% for
221 Smallmouth Bass (Figure S9). Likewise, performances of all methods in reproducing the rates
222 within a factor of two were comparatively similar for all species (Figure 4). When a more robust
223 dataset (cases with a minimum time span of 15 years and 10 sampling years) was considered, all
224 methods resulted in rates that were within a factor of three in 97-100% of the cases for Lake
225 Trout, Northern Pike and Walleye (Figure 4). The performance of the composite samples in
226 reproducing the rates of change for Smallmouth Bass was less (86-90%) compared to the other
227 three species (Figure 4), indicating that Smallmouth Bass is the least preferred species for trend
228 monitoring when a composite method is utilised.

229 As expected, the composite methods that resulted in fewer reductions in the number of
230 fish mercury measurements provided better estimates of the rates of changes in the fish mercury
231 levels (Figures 2 and 4). Although reductions (55/60%) in number of measurements from
232 method 6 were comparable to method 2 (54/64%), method 6 provided better estimates of the
233 rates of change (Figure 2 and 4). Furthermore, the performance of method 6 was comparable to
234 the method 3, which consisted of relatively more mercury measurements (Figures 2 and 4). The
235 differences in the performance of the methods in reproducing the rates were minimal when cases
236 with a minimum time span of 15 years and 10 sampling years were considered (Figure 4).

237

238 **4. Discussion**

239 Composite sampling combines environmental samples or subsamples to form a new
240 sample on which chemical or biological analyses are performed. Compared to evaluating
241 individuals, composite sampling is beneficial as it decreases analytical cost by analyzing fewer

242 samples and reduces/simplifies the sample handling process (USEPA, 2002). Composite
243 sampling is recommended when laboratory costs are substantially greater than field sampling
244 costs (USEPA, 2002). The collection of a few more fish samples at a particular location may not
245 substantially increase the field cost. However, the analytical savings associated with composite
246 sampling in long-term fish mercury monitoring and for issuance of fish consumption advisories
247 can be substantial, especially over time. For example, the approximately 60% reductions in
248 sample analyses in the OMOECC dataset used in this study would have resulted in
249 approximately 134,000 fewer fish mercury analyses over the 47 year period, which sums to
250 about \$5,400,000 (or \$114,000 per year) at an average rate of \$40 per sample. Similarly, about
251 \$1,000,000 could be saved for the dataset compiled by USGS from data collected by US states
252 (Hearn et al., 2006). Further, the composite sampling would have resulted in substantial saving in
253 other operational costs due to reduced number of samples to handle. Although the extent of cost
254 saving would depend on nature of the program (e.g., how many individual samples of which fish
255 species and sizes are presently analysed for mercury) and analytical cost, which has been
256 declining with advances in the analytical technology, the results presented in this study show that
257 savings can be achieved without any major impact on the quality of the advisories or temporal
258 trend assessments.

259 There are, however, some potential disadvantages of the composite sampling approach.
260 For example, composite sampling can result in a loss of information on extreme contamination
261 levels and variability. Although this is true in many cases, a composite method retaining one or
262 two largest and smallest individual samples as suggested in this study can potentially capture
263 extreme fish mercury levels due to the strong relationship of fish size and mercury concentration.
264 Although method 6 considered in this study may not preserve individual samples, a power series

265 relationship between fish length and mercury indicates that compositing within a 5 cm fish size
266 bin would likely be able to provide values closer to the extreme levels. This could be a result of
267 the pattern in fish mercury levels, where even though there is a strong relationship between fish
268 length and mercury levels, it is not necessary that the biggest fish has the highest concentration
269 and the smallest fish has the lowest concentration likely due to differences in mercury levels in
270 spatially integrated fish samples. Compositing reduces sample size, and as such decreases
271 statistical power; however, statistical formulas can be used to derive composite size that results
272 in a sufficient power (Rohlf et al., 1996). The composite methods examined in this study also
273 resulted in some loss of statistical significance (Figure S6). Nevertheless, the methods
274 performed reasonably well in reproducing the advisories and temporal trends (Figures 3, 4, S7).

275 If contaminants other than mercury are also of interest, further evaluation of the
276 compositing methods may be necessary. For North America, other major contaminants of
277 concerns include persistent organic pollutants (POPs) for which compositing is often performed
278 (Hites et al., 2004; Gewurtz et al., 2011a) for studies focused on the health of fish themselves
279 and not on the generation of fish consumption advice. Gewurtz et al. (2011a) found compositing
280 fish samples appropriate for temporal trend monitoring of polychlorinated biphenyls (PCBs)
281 based on a limited evaluation of Lake Ontario lake trout measurements from different Canadian
282 and U.S. monitoring programs. However, their evaluation did not consider the impact of
283 compositing on the ability to detect outliers. It should be noted that the relationship between fish
284 length and POPs, such as PCBs, is much weaker than is typically observed for mercury (e.g.,
285 Gewurtz et al., 2011b). As such, compositing fish samples based on size categories (e.g., method
286 6 in this study) may be less effective in capturing outliers for POPs. However, many agencies use
287 the "75% rule" (i.e., the length of the smallest fish in a composite should be at least 75% of the

288 length of the largest fish) for compositing fish samples for POP monitoring (e.g., Stahl et al.,
289 2009). The method 6 considered in this study will composite samples within a 5 cm size range
290 (Figure 1a,g) and follow the 75% rule (except for fish smaller than 15 cm, which are generally
291 not considered sport fish anyway). Similarly, the method 2 (and probably the other methods
292 considered) will also create composites (Figure 1a,c) that has a high potential to follow the 75%
293 rule (Tables S11-S12), depending on the extent of sample collection by a program. As such, the
294 compositing methods and findings of this study may also be suitable for monitoring POPs in fish.

295 A reliable temporal trend analysis depends on within-year samples and duration of
296 monitoring (Sokal and Rohlf, 1995). Based on an exploratory analysis performed on data
297 collected by some Great Lakes biomonitoring programs and a comparison with the literature, it
298 was concluded that >10 years of monitoring with 10-15 samples per year is optimal to achieve
299 80% statistical power, which is typically considered adequate (Gewurtz et al., 2011a). This is
300 largely due to diminished sensitivity of a temporal trend analysis to start and end points when a
301 reasonable length of monitoring data is available (Gewurtz et al., 2011a). In this study, the
302 correspondence between the results from the regular and composite methods improved when a
303 longer time span and increased number of sampling years were considered (Figures 4, S7-S9).
304 As such, compositing samples may not be advisable for a short term assessment; however, the
305 accuracy of the regular method based on individual samples may also be poor.

306 In this study, we utilized skinless, boneless fillet mercury measurements to evaluate the
307 compositing methods. However, some monitoring programs use muscle plug or whole fish
308 measurements to track environmental conditions. Since fish fillet, muscle plug and whole fish
309 mercury measurements can be linked to one another (Baker et al., 2004; Peterson et al., 2005),
310 findings from this study should be applicable to muscle plug and whole fish mercury monitoring

311 studies as well. The Ontario’s fish contaminant monitoring is conducted exclusively in temperate
312 environments and thus the results from this study have broad applicability to other monitoring
313 programs in temperate latitudes. Although the in-depth analyses conducted on an extensive
314 dataset indicate that the findings should be applicable to tropical environment as well, further
315 work to verify these results in tropical environment may be warranted.

316 In summary, we explored the suitability of six composite methods for fish mercury
317 monitoring using an extensive dataset. The methods resulted in varying amount of reductions in
318 number of samples to be analyzed. In general, all compositing methods performed well for both
319 advisories on consumption of fish and temporal trend monitoring. The methods resulting in
320 lower reductions in sample count performed marginally better. Overall, compositing samples
321 would have resulted in a substantial cost savings for OMOECC (approximately \$5.4 M over 47
322 years assuming 60% sample reduction) , and should be considered in fish mercury monitoring
323 especially in long-term extensive monitoring programs or when study cost is a concern.

324

325 **Acknowledgements**

326 We thank Ontario Ministry of the Environment and Climate Change, Canada, for the long-term
327 fish mercury dataset.

328

329 **Supplementary Material**

330 Additional 12 tables and 9 figures. This material is available free of charge via the
331 Internet at ??????

332

333 **Notes** The authors declare no competing financial interest.

334 **References**

335

336 Azim, M., Kumarappah, A., Bhavsar, S., Backus, S., Arhonditsis, G., 2011. Detection of the
337 spatiotemporal trends of mercury in Lake Erie fish communities: A Bayesian approach.
338 Environ. Sci. Technol. 45, 2217-2226.

339 Baker, R.F., Blanchfield, P.J., Paterson, M.J., Flett, R.J., Wesson, L., 2004. Evaluation of
340 nonlethal methods for the analysis of mercury in fish tissue. T. Am. Fish. Soc. 133, 568-
341 576.

342 Bhavsar, S.P., Gewurtz, S.B., McGoldrick, D.J., Keir, M.J., Backus, S.M., 2010. Changes in
343 mercury levels in Great Lakes fish between 1970s and 2007. Environ. Sci. Technol. 44,
344 3273-9.

345 Blomqvist, P., 2001. A proposed standard method for composite sampling of water chemistry
346 and plankton in small lakes. Environ. Ecol. Stat. 8, 121-134.

347 Braune, B.M., Noble, D.G., 2009. Environmental contaminants in Canadian shorebirds. Environ.
348 Monit. Assess. 148, 185-204.

349 Carlson, D.L., Swackhamer, D.L., 2006. Results from the US Great Lakes fish monitoring
350 program and effects of lake processes on bioaccumulative contaminant concentrations. J.
351 Great Lakes Res. 32, 370-385.

352 Depew, D.C., Burgess, N.M., Anderson, M.R., Baker, R., Bhavsar, S.P., Bodaly, R.A., Eckley,
353 C.S., Evans, M.S., Gantner, N., Graydon, J.A., Jacobs, K., LeBlanc, J.E., St Louis, V.L.,
354 Campbell, L.M., 2013. An overview of mercury concentrations in freshwater fish species:
355 a national fish mercury dataset for Canada. Can. J. Fish. Aquat. Sci. 70, 436-451.

356 Evers, D.C., DiGangi, J., Petrlík, J., Buck, D.G., Šamánek, J., Beeler, B., Turnquist, M.A.,
357 Hatch, S.K., Regan, K., 2013. Global mercury hotspots: New evidence reveals mercury
358 contamination regularly exceeds health advisory levels in humans and fish
359 worldwide Biodiversity Research Institute, Gorham, Maine; IPEN, Göteborg, Sweden p.
360 20.

361 French, T.D., Petro, S., Reiner, E.J., Bhavsar, S.P., Jackson, D.A., 2011. Thirty-year time series
362 of PCB concentrations in a small invertivorous fish (*Notropis hudsonius*): An
363 examination of post-1990 trajectory shifts in the lower Great Lakes. *Ecosystems* 14, 415-
364 429.

365 Gewurtz, S.B., Backus, S.M., Bhavsar, S.P., McGoldrick, D.J., de Solla, S.R., Murphy, E.W.,
366 2011a. Contaminant biomonitoring programs in the Great Lakes region: Review of
367 approaches and critical factors. *Environ. Rev.* 19, 162-184.

368 Gewurtz, S.B., Bhavsar, S.P., Fletcher, R., 2011b. Influence of fish size and sex on mercury/PCB
369 concentration: Importance for fish consumption advisories. *Environ. Int.* 32, 425-434.

370 Greenfield, B.K., Melwani, A.R., Allen, R.M., Slotton, D.G., Ayers, S.M., Harrold, K.H.,
371 Ridolfi, K., Jahn, A., Grenier, J.L., Sandheinrich, M.B., 2013. Seasonal and annual trends
372 in forage fish mercury concentrations, San Francisco Bay. *Sci. Total Environ.* 444, 591-
373 601.

374 Hearn, P.P., Wente, S.P., Donato, D.I., Aguinaldo, J.J., 2006. EMMMA: A Web-based System
375 for Environmental Mercury Mapping, Modeling, and Analysis. U.S. Geological Survey
376 Open File Report 2006-1086, 13p.

377 Hites, R.A., Foran, J.A., Carpenter, D.O., Hamilton, M.C., Knuth, B.A., Schwager, S.J., 2004.
378 Global assessment of organic contaminants in farmed salmon. *Science* 303, 226-9.

379 OMOECC, 2015. 2015-2016 Guide to Eating Ontario Fish. Ontario Ministry of the Environment
380 and Climate Change, Toronto, Ontario, Canada.

381 Pantazopoulos, P., Sawyer, J.M., Turyk, M.E., Diamond, M., Bhavsar, S.P., Mergler, D.,
382 Schantz, S., Ratnayake, N., Carpenter, D.O., 2013. Fatty acids in Great Lakes lake trout
383 and whitefish. *J. Great Lakes Res.* 39, 120-7.

384 Peterson, S.A., Van Sickle, J., Hughes, R.M., Schacher, J.A., Echols, S.F., 2005. A biopsy
385 procedure for determining filet and predicting whole-fish mercury concentration. *Arch.*
386 *Environ. Con. Tox.* 48, 99-107.

387 R Core Development Team, 2015. R: A language and environment for statistical computing. R
388 Foundation for Statistical Computing, Vienna, Austria. .

389 Rajagopal, R., Williams, L.R., 1989. Economics of sample compositing as a screening tool in
390 ground-water quality monitoring. *Ground Water Monit. R.* 9, 186-192.

391 Rohlf, F.J., Akcakaya, H.R., Ferraro, S.P., 1996. Optimizing composite sampling protocols.
392 *Environ. Sci. Technol.* 30, 2899-2905.

393 Sokal, R.R., Rohlf, F.J., 1995. *Biometry: the principles and practice of statistics in biological*
394 *research*, 3rd Edition ed. W. H. Freeman and Co., New York, NY, USA.

395 Stahl, L., Snyder, B., Olsen, A., Pitt, J., 2009. Contaminants in fish tissue from US lakes and
396 reservoirs: a national probabilistic study. *Environ. Monit. Assess.* 150, 3-19.

397 Stern, G.A., Macdonald, R.W., Outridge, P.M., Wilson, S., Chetelat, J., Cole, A., Hintelmann,
398 H., Loseto, L.L., Steffen, A., Wang, F., Zdanowicz, C., 2012. How does climate change
399 influence arctic mercury? *Sci. Total Environ.* 414, 22-42.

400 Turle, R., Collins, B., 1992. Validation of the use of pooled samples for monitoring of
401 contaminants in wildlife. *Chemosphere* 25, 463-9.

402 UNEP, 2013a. Global Mercury Assessment 2013: Sources, Emissions, Releases and
403 Environmental Transport. UNEP Chemical Branch, UNEP Chemical Branch, Geneva,
404 Switzerland.

405 UNEP, 2013b. Minamata Convention on Mercury. <http://www.mercuryconvention.org>.
406 UNEP/WHO, 2008. Guidance for identifying populations at risk from mercury exposure.
407 Geneva, Switzerland, p. 176.

408 USEPA, 2002. Guidance on choosing a sampling design for environmental data collection US
409 EPA, Washington, DC, p. 166.

410 USEPA, 2013a. Fish Consumption Advisories. USEPA. <http://www2.epa.gov/fish-tech>.
411 USEPA, 2013b. Guidance for Assessing Chemical Contaminant Data for Use In Fish Advisories.
412 USEPA. EPA 823-B-00-007.
413 <http://water.epa.gov/scitech/swguidance/fishshellfish/techguidance>.
414
415

Figure 1: Illustration of six compositing methods considered in the study.

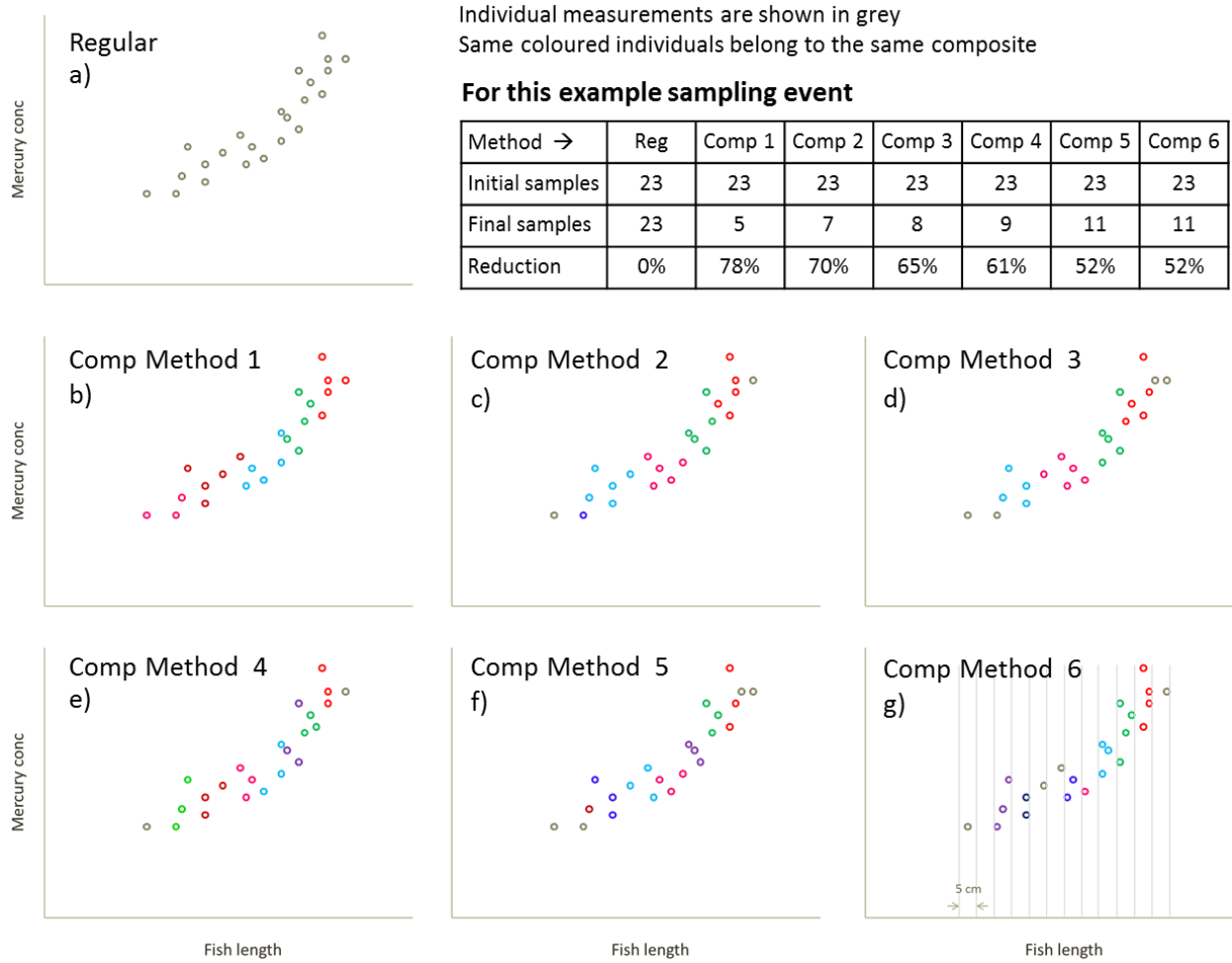


Figure 2: Overall reduction (%) in number of samples per sampling event (location/year/species) analyzed in each of the six composite methods compared to the regular method of analyzing all individual fish samples for mercury.

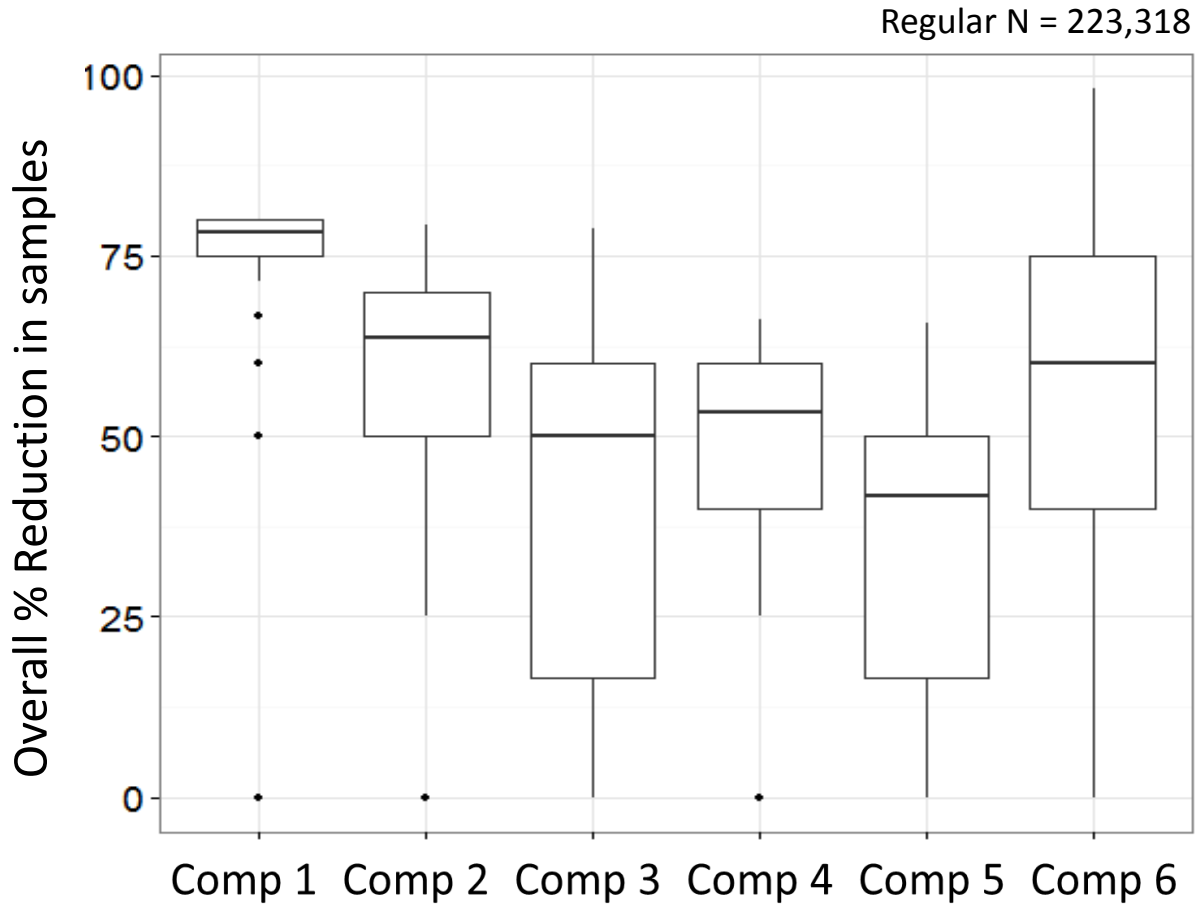


Figure 3: Comparison on fish consumption advisories for mercury for the general and sensitive populations using composite methods compared to the current OMOECC method of analyzing individual fish samples.

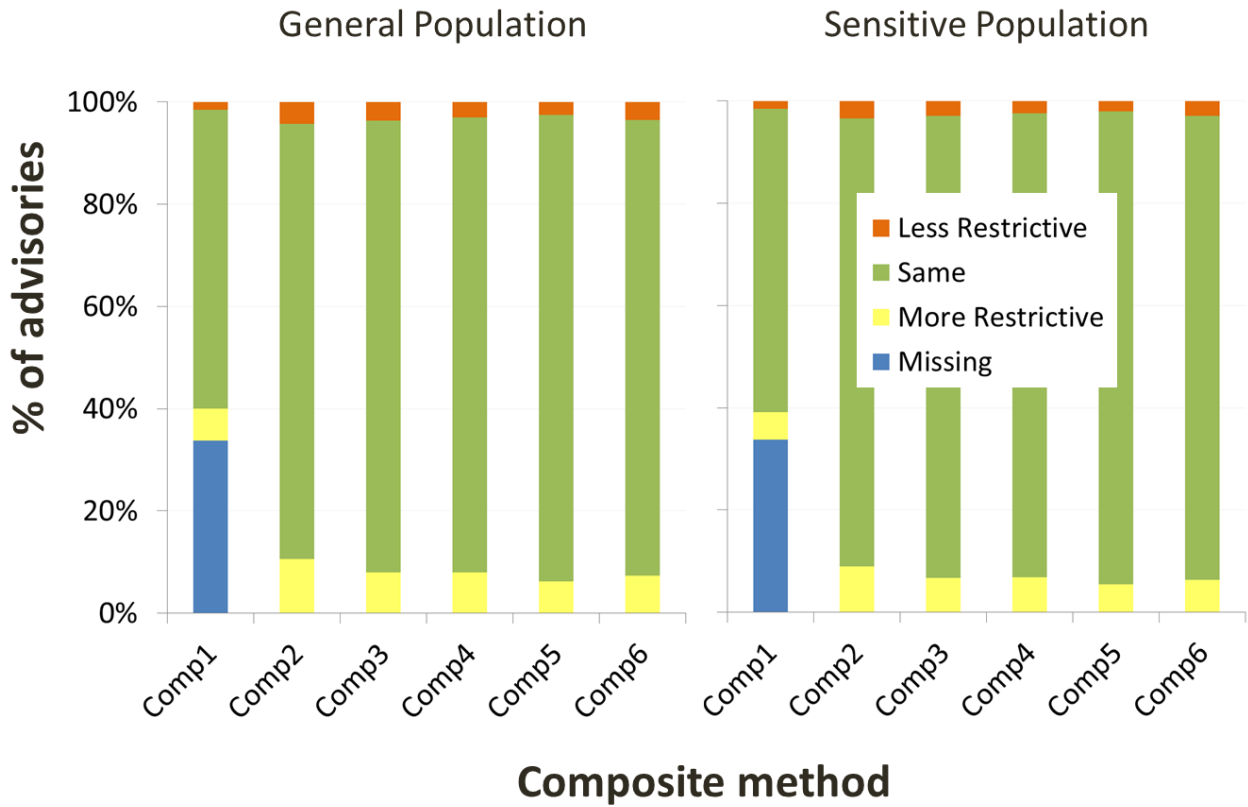


Figure 4: Comparison of rates of change in fish mercury levels of the six composite methods with those from the current OMOECC method of analyzing individual fish samples for mercury. The results have been presented as percentage of the total number of rate estimates within 2 and 3 times the corresponding rates from the current OMOECC method.

